

# Homogeneous Electro-Fenton Oxidative Degradation of Reactive Brilliant Blue Using a Graphene Doped Gas-Diffusion Cathode

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A novel graphene doped gas diffusion electrode (GE/graphite-PTFE GDE) was prepared in this work with graphite powder and PTFE as starting materials. An undivided electro-Fenton system with this self-made GDE as cathode was fabricated to evaluate the activity of GE/graphite-PTFE GDE for H<sub>2</sub>O<sub>2</sub> electro-generation. Compared with bare graphite-PTFE GDE, the electro-generated H<sub>2</sub>O<sub>2</sub> concentration on GE/graphite-PTFE GDE was improved three times and can be reached to ca. 187 mg/L after 3 h electrolysis. The maximum production of H<sub>2</sub>O<sub>2</sub> was obtained at pH 3.0, the current density of 2.0 mA/cm<sup>2</sup> and M<sub>graphite</sub>:M<sub>GE</sub>:M<sub>PTFE</sub> = 8:1:2. The influences of operating parameters including solution pH and Fe<sup>2+</sup> concentration were investigated on Reactive brilliant blue (KN-R) removal by the proposed electro-Fenton oxidation process. Under the optimal condition, nearly 80% of KN-R and 33.3% total organic carbon (TOC) could be removed after 180 min electrolysis, indicating that the GE/graphite/PTFE GDE was effective in H<sub>2</sub>O<sub>2</sub> electro-generation and the proposed homogeneous electro-Fenton system was effective in KN-R oxidative degradation.

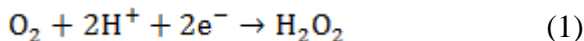
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**Keywords:** Graphene; Gas diffusion cathode; Electro-Fenton; H<sub>2</sub>O<sub>2</sub>; Reactive Brilliant Blue

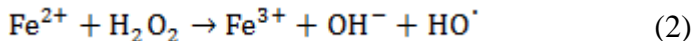
## 1. INTRODUCTION

Wastewater generated by textile industries contains considerable amounts of dyes. It is well known that some dyes and degradation products such as aromatic amines are highly carcinogenic [1]. It is a challenge to remove dyes from wastewater because the traditional treatment processes have been proved to be insufficient [2-4]. Recently, electro-Fenton methods, which combine with ferrous ion (Fe<sup>2+</sup>) addition and the in-situ electrogenerated H<sub>2</sub>O<sub>2</sub> have attracted much attention, since they have been widely used to treat various dyes wastewater [5-8]. Advantages of electro-Fenton are numerous,

including ease of control, amenability to automation, high efficiency, and environmental compatibility, etc [9]. In electro-Fenton process,  $\text{H}_2\text{O}_2$  can be produced by a two-electron reduction of oxygen (equation (1)) at appropriate cathodic potential on certain electrodes such as reticulated vitreous carbon, graphite and gas diffusion electrode (GDE) [5,6,10,11].



Hydroxyl radicals ( $\bullet\text{OH}$ ) are generated in the solution by the well-known Fenton's reaction between  $\text{Fe}^{2+}$  and electrogenerated  $\text{H}_2\text{O}_2$  (equation (2)).

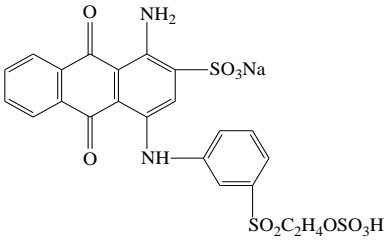


Cathode materials, which have high  $\text{H}_2\text{O}_2$  production ability, are very important for the oxidation ability of electro-Fenton system. Recently, cathode materials, such as carbon/polytetrafluoroethylene (C/PTFE) GDE [5,6,10,11], graphite [12-14], carbon nanotubes [15], carbon felt [16-19], and reticulated vitreous carbon [20,21] are used as electrode materials for the electrochemical production of  $\text{H}_2\text{O}_2$ . Graphene (GE) has emerged as an interesting material because of its particular nanostructure, outstanding electron-transfer capacity, good stability and large accessible surface area [22,23]. Considering its remarkable properties and its relatively easy preparation approach, the novel GDE was prepared with GE, graphite powder and PTFE in this work. The electroactive surface area of this self-made GDE was characterized by the electrochemical methods. The activity, current efficiency and the stability of this self-made GDE for the electrogeneration of  $\text{H}_2\text{O}_2$  was investigated in detail. Reactive brilliant blue (KN-R) was used as a model anthraquinone dye and was degraded by the electro-Fenton process in an undivided cell. The influence of various parameters on KN-R removal such as pH and concentration of  $\text{Fe}^{2+}$  were systematically evaluated.

## 2. EXPERIMENTAL

### 2.1. Reagent and material

**Table 1.** Physicochemical properties of Reactive Brilliant Blue (KN-R)

Chemical name	Reactive Brilliant Blue (KN-R)
Chemical formula	$\text{C}_{22}\text{H}_{16}\text{N}_2\text{Na}_2\text{O}_{11}\text{S}_3$
Chemical structural	
Molecular Weight	626.56
Class	Anthraquinone
Maximum absorption wavelength	590 nm

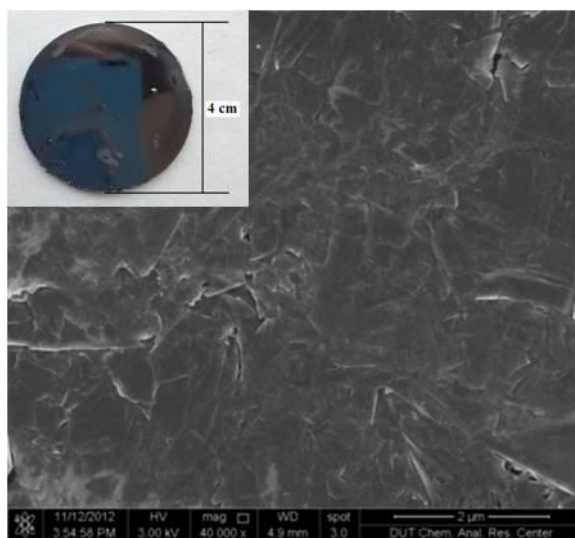
Graphite powder (600 mesh, analytical grade), PTFE (60%) and stainless steel mesh (20 mesh) were purchased from Beijing Chemical Regent Plant (China).  $\text{NaNO}_3$ ,  $\text{KMnO}_4$  and  $\text{H}_2\text{O}_2$  (30 wt%)

were of analytical grade and obtained from Shanghai Chemical Reagent Co. (China). Reactive Brilliant Blue (KN-R, analytical grade) was purchased from Tianjing Kemi-ou Chemical Regent Co. (China) and its physicochemical properties are shown in Table 1.

All other chemicals including  $\text{H}_2\text{SO}_4$ , anhydrous  $\text{Na}_2\text{SO}_4$ ,  $\text{NaOH}$ ,  $\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$ ,  $\text{TiOSO}_4 \cdot 2\text{H}_2\text{O}$  and anhydrous ethanol were of analytical grade. Deionized water was used in all the experiments.

## 2.2. Fabrication of GDE

Graphene (GE) was obtained by chemical oxidated graphene oxide, which was synthesized from natural graphite powder using a modified Hummers' method [24,25]. Graphite powder, GE and PTFE solution with the mass ratio of 100:1:2, 50:1:2 and 8:1:2 were added and mixed respectively in an ultrasonic bath for 15 min at room temperature. The mixture was stirred at about  $80^\circ\text{C}$  until it became ointment, and then was put into the cylinder mold and formed a circle sheet (4 cm diameter, 1.5 mm thickness) by pressed at 15 MPa for 2 min. The fabricated GE/graphite-PTFE GDEs were finally obtained by calcinations at  $300^\circ\text{C}$  for 2 h (Fig. 1).



**Figure 1.** Photograph (inset) and SEM image of GE/graphite-PTFE GDE with mass ratio of  $M_{\text{graphite}}:M_{\text{GE}}:M_{\text{PTFE}} = 8:1:2$

For comparison, graphite-PTFE GDE was also prepared using the same method except that no GE was added.

## 2.3. Electrolysis experiments

The in-situ electrogeneration of  $\text{H}_2\text{O}_2$  were carried out with galvanostatic electrolysis method in a three-electrode undivided cell (200 mL) at room temperature, which was controlled by REX DJS-292 potentiostat/galvanostat (Shanghai Cany Precision Instrument Co., Ltd). The prepared

GE/graphite-PTFE GDE (12 cm<sup>2</sup> area) was selected as working electrode, a Pt sheet (6 cm<sup>2</sup> area) and saturated calomel electrode (SCE) as counter and reference electrodes, respectively. The distance between the working electrode and counter electrode was 3 cm. Prior to electrolysis, 200 mL Na<sub>2</sub>SO<sub>4</sub> aqueous solution (0.05 mol/L) was added in the undivided cell. The solution was adjusted to the desired pH value (from 1.5 to 8.0) using H<sub>2</sub>SO<sub>4</sub> and NaOH and vigorously stirred with a magnetic bar to enhance the mass transport. The solution pH was measured with a Metrohm 654 pH meter (Switzerland). In order to reach high dissolved oxygen concentration for high H<sub>2</sub>O<sub>2</sub> production, pure oxygen at 0.02 m<sup>3</sup> h<sup>-1</sup> was insufflated into the cell through a porous pipe-diffuser placed right under the cathode. The initial KN-R concentration was 80 mg/L corresponding to 18 mg/L total organic carbon (TOC). The current density changed from 0.5 to 3.0 mA/cm<sup>2</sup> and the concentration of Fe<sup>2+</sup> changed from 0.1 mmol/L to 1.0 mmol/L. For comparison, using graphite-PTFE GDE as working electrode repeated the above-mentioned experiment in order to evaluate the differences in the electrogeneration of H<sub>2</sub>O<sub>2</sub> and dye removal efficiency.

#### 2.4. Analytical methods

All samples extracted from the treated solutions were filtered through 25 μm pore size filter paper before analysis. H<sub>2</sub>O<sub>2</sub> concentration was analyzed spectrophotometrically (pharmaspec, UV-1700, Japan) by measuring the absorbance at λ = 408 nm (ε = 1010 M<sup>-1</sup> cm<sup>-1</sup>). The cell electric potential was recorded during electrolysis.

The current efficiency (CE) for H<sub>2</sub>O<sub>2</sub> formation in the absence of Fe<sup>2+</sup> is determined as:

$$CE = \frac{nF[H_2O_2]V}{Q_t} \times 100\% \quad (3)$$

Where  $n$  ( $n = 2$ ) is the number of electrons involved in the electrochemical reaction (1),  $F$  is the Faraday constant (96,485 C/mol),  $[H_2O_2]$  represents the concentration of H<sub>2</sub>O<sub>2</sub> in the reaction solution (mol/L),  $V$  is the total volume of electrolytic solution (L), and  $Q_t$  is the cumulative charge applied to the system by any given time (C), obtained by integrating the current. TOC analyses were carried out using a Shimadzu 5050 TOC-V<sub>CPH</sub> analyzer.

After the homogeneous electro-Fenton oxidation, KN-R concentration was also analyzed spectrophotometrically by measuring the absorbance at λ = 580 nm.

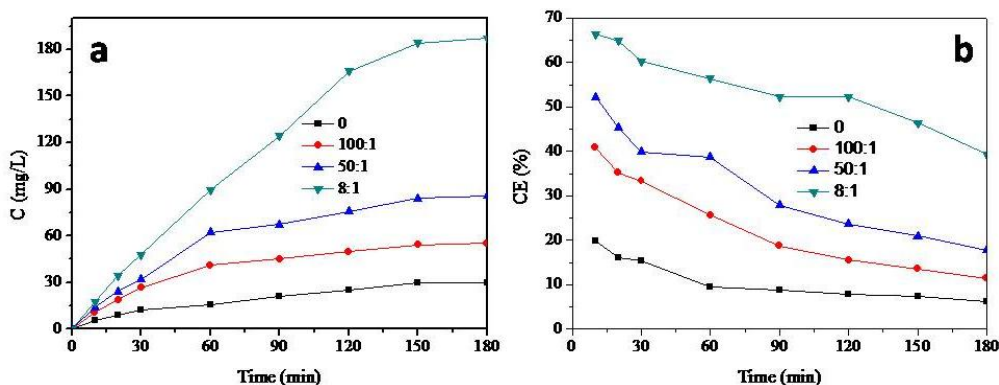
### 3. RESULT AND DISCUSSION

#### 3.1. Electro-generation of H<sub>2</sub>O<sub>2</sub>

##### 3.1.1. Comparison of electro-generated H<sub>2</sub>O<sub>2</sub> on different GDEs

In order to investigate the electro-generation of H<sub>2</sub>O<sub>2</sub> on GE/graphite-PTFE and graphite-PTFE GDE, several experiments were performed in the undivided electrolysis system containing 0.05 mol/L Na<sub>2</sub>SO<sub>4</sub> (pH 3) at current density of 2.0 mA/cm<sup>2</sup>. Fig. 2a shows the change of H<sub>2</sub>O<sub>2</sub> concentration with the electrolysis time on different GDEs. With the increase of the electrolysis time, the concentration of

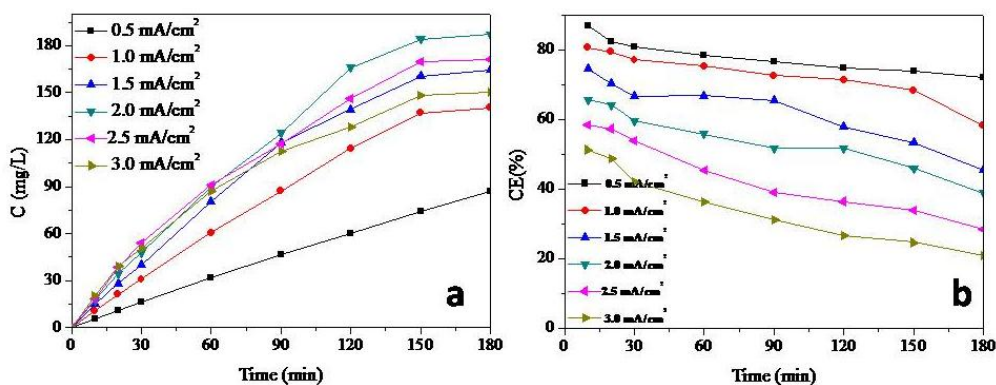
H<sub>2</sub>O<sub>2</sub> all increased in two systems. After 180 min electrolysis, the concentration of H<sub>2</sub>O<sub>2</sub> produced on GE/graphite-PTFE GDE (M<sub>graphite</sub>:M<sub>GE</sub>:M<sub>PTFE</sub> = 8:1:2) was 187.1 mg/L, and it was nearly four times higher than that of graphite-PTFE electrode (47.6 mg/L, M<sub>graphite</sub>:M<sub>PTFE</sub> = 8:2).



**Figure 2.** Changes in the electro-generated H<sub>2</sub>O<sub>2</sub> concentration (a) and current efficiency (b) with electrolysis time on GE/graphite-PTFE GDE with various mass ratio of M<sub>graphite</sub>:M<sub>GE</sub>:M<sub>PTFE</sub>. Conditions: O<sub>2</sub> flow rate of 0.02 m<sup>3</sup>/h, 0.05 mol/L Na<sub>2</sub>SO<sub>4</sub>, pH 3.0, current density of 2.0 mA/cm<sup>2</sup>.

The changes in current efficiency showed in Fig. 2b exhibits the same trend as that in the accumulated concentration of H<sub>2</sub>O<sub>2</sub>. It can be attributed to the faster electro transfer accelerated via the introduction of GE in the graphite-PTFE GDE.

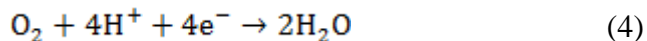
3.1.2. Effect of current density on the electro-generation of H<sub>2</sub>O<sub>2</sub>



**Figure 3.** Influence of applied current density on the electro-generated H<sub>2</sub>O<sub>2</sub> concentration (a) and current efficiency (b) on GE/graphite-PTFE GDE. Conditions: O<sub>2</sub> flow rate of 0.02 m<sup>3</sup>/h, 0.05 mol/L Na<sub>2</sub>SO<sub>4</sub>, pH 3.0, M<sub>graphite</sub>:M<sub>GE</sub>:M<sub>PTFE</sub> = 8:1:2.

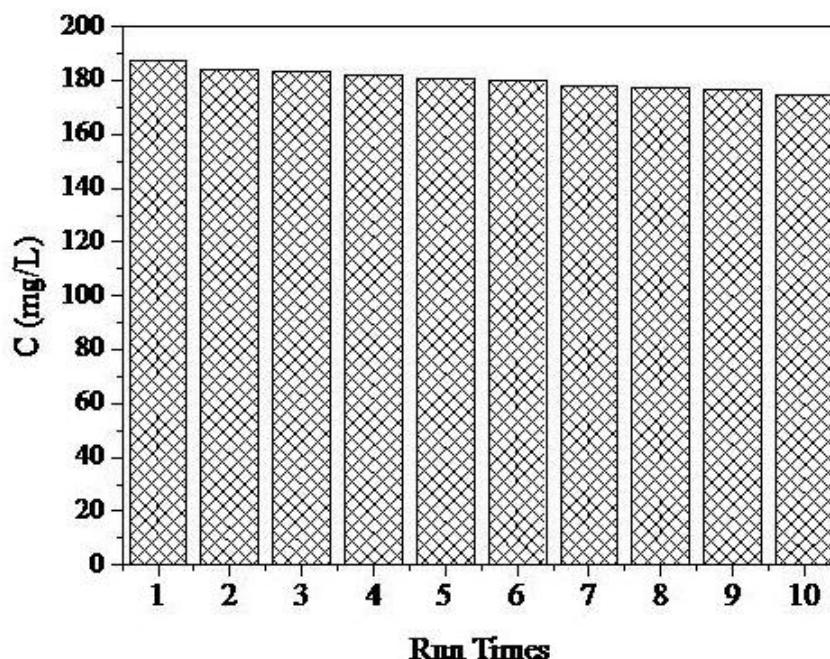
Considering that current density had notable influence on the electro-generation of H<sub>2</sub>O<sub>2</sub>, the electrolysis experiments were also carried out using various current densities ranging from 0.5 to 3.0 mA/cm<sup>2</sup> on GE/graphite-PTFE GDE (M<sub>graphite</sub>:M<sub>GE</sub>:M<sub>PTFE</sub> = 8:1:2). As shown in Fig. 3a, the

concentration of  $\text{H}_2\text{O}_2$  increases as current density increasing from 0.5 to 2.0  $\text{mA}/\text{cm}^2$ , and then rapidly decreases with the further increase of current density. These results can be explained by the change in cell potential. The average cell potential measured was ca. 1.44, 1.82, 2.47, 3.30, 4.45, 5.37 and 5.89 V during the electrolysis for the applied current density of 0.5, 1.0, 1.5, 2.0, 2.5 and 3.0  $\text{mA}/\text{cm}^2$ , respectively. It is reported that at the cell potential values higher than 4.45 V, the reduction of  $\text{O}_2$  leads to the production of  $\text{H}_2\text{O}$  instead of  $\text{H}_2\text{O}_2$  through the equation (4) [26], resulting also in the decrease in current efficiency when the applied current density was above 2.0  $\text{mA}/\text{cm}^2$  (Fig. 3b).



### 3.1.3. Stability for $\text{H}_2\text{O}_2$ electro-generation

Repeated experiments were performed under the conditions of pH of 3.0, current density of 2.0  $\text{mA}/\text{cm}^2$  and  $M_{\text{graphite}}:M_{\text{GE}}:M_{\text{PTFE}} = 8:1:2$  to examine the reusability of the prepared GE/graphite-PTFE cathode in producing  $\text{H}_2\text{O}_2$ .

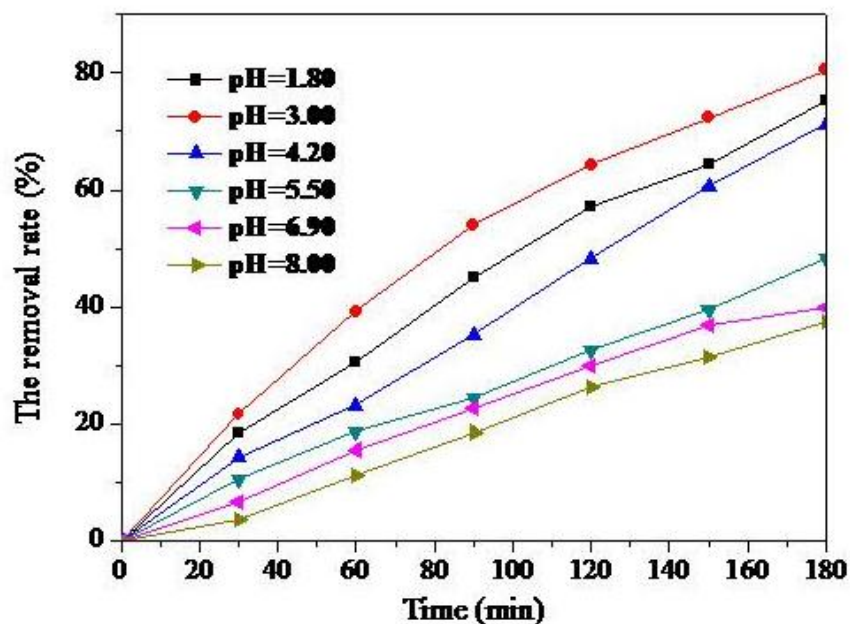


**Figure 4.** Stability of the GE/graphite-PTFE GDE for the electro-generation of  $\text{H}_2\text{O}_2$  after 180 min electrolysis. Conditions:  $\text{O}_2$  flow rate of 0.02  $\text{m}^3/\text{h}$ , 0.05 mol/L  $\text{Na}_2\text{SO}_4$ , pH 3.0, current density of 2.0  $\text{mA}/\text{cm}^2$ ,  $M_{\text{graphite}}:M_{\text{GE}}:M_{\text{PTFE}} = 8:1:2$ .

As shown in Fig. 4, the concentration of electro-generated  $\text{H}_2\text{O}_2$  declined less 7% from 187.7 to 174.6 mg/L after ten times reuse, indicating the excellent stability of GE/graphite-PTFE GDE for  $\text{H}_2\text{O}_2$  electro-generation.

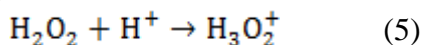
### 3.2. Degradation of KN-R

#### 3.2.1. Effect of solution pH

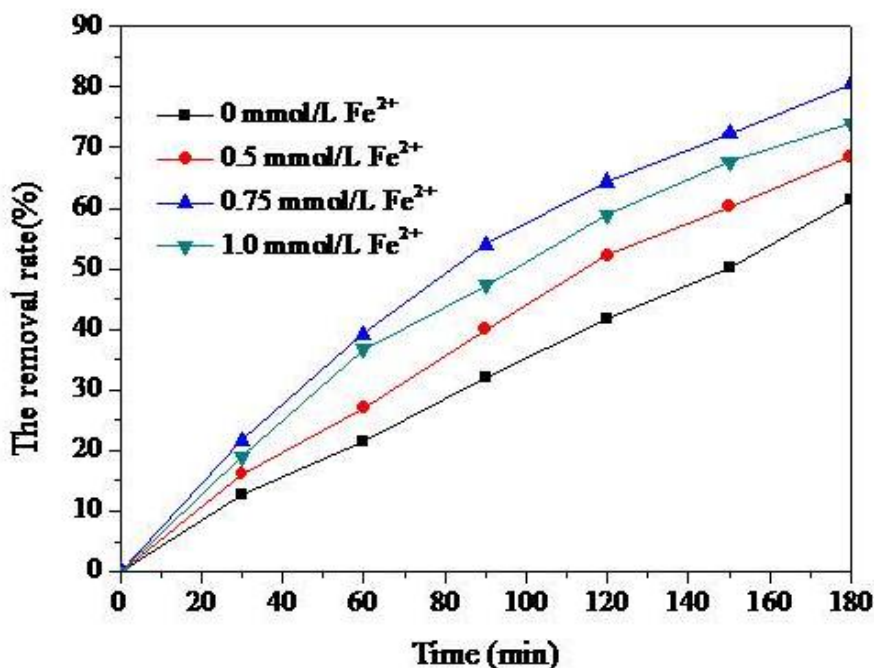


**Figure 5.** Influence of solution pH on KN-R degradation by homogeneous electro-Fenton process using GE/graphite-PTFE GDE. Conditions:  $O_2$  flow rate of  $0.02 \text{ m}^3/\text{h}$ ,  $0.05 \text{ mol/L Na}_2\text{SO}_4$ , current density of  $2.0 \text{ mA/cm}^2$ ,  $0.75 \text{ mmol/L Fe}^{2+}$  and  $M_{\text{graphite}}:M_{\text{GE}}:M_{\text{PTFE}} = 8:1:2$ .

It was well known that the solution pH was an important parameter for Fenton oxidation process. Fig. 5 presents KN-R decay in various pH solutions (from pH 1.5 to 8.0) containing  $0.75 \text{ mM Fe}^{2+}$  using GE/graphite-PTFE GDE. As can be seen, electro-Fenton process is effective and the highest oxidation activity is achieved at pH 3.0. About 80% KN-R was removed after 180 min electrolysis, which was entirely consistent with the reported by other authors for the degradation of organic pollutants via the electro-Fenton methods and it is also very close to the optimum pH of 2.8 for Fenton's reaction [27]. At low pH values below 3, the electro-generated  $H_2O_2$  will solvate a proton to form an oxonium ion ( $H_3O_2^+$ , as shown in equation (5)) which enhanced its stability and lost its activity in Fenton chain reaction.

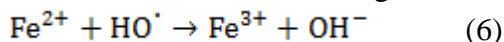


When the solution pH was above 3.0, the coagulation accompanying with the oxidation simultaneously occurred, involving the formation of hydroxyl-complexes of iron and the polymerization reaction of the subsequent products [28]. The formation of these ferric hydroxide precipitation not only decreases  $Fe^{2+}$  regeneration, but also inhibits  $H_2O_2$  formation by partially coating the GE/graphite-PTFE GDE surface, thus reducing electro-Fenton oxidation efficiency. An inspection of the GE/graphite-PTFE GDE conducted in  $pH > 5.5$  disclosed the occurrence of yellow  $Fe(OH)_3$  precipitation on its surface.

3.2.2. Effect of  $\text{Fe}^{2+}$  concentration

**Figure 6.** Influence of  $\text{Fe}^{2+}$  concentration on KN-R degradation by homogeneous electro-Fenton process using GE/graphite-PTFE GDE. Conditions:  $\text{O}_2$  flow rate of  $0.02 \text{ m}^3/\text{h}$ ,  $0.05 \text{ mol/L}$   $\text{Na}_2\text{SO}_4$ , current density of  $2.0 \text{ mA/cm}^2$ , pH 3.0 and  $M_{\text{graphite}}:M_{\text{GE}}:M_{\text{PTFE}} = 8:1:2$ .

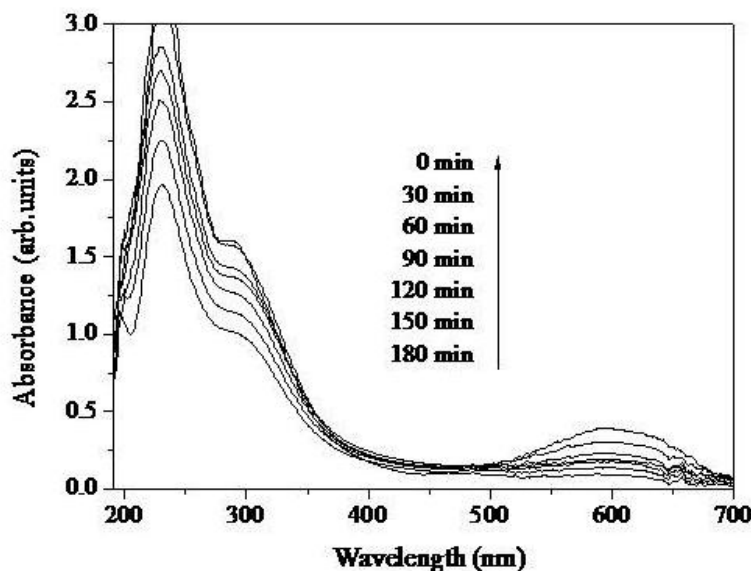
Organics degradation by electro-Fenton process not only depended on electrogenerated  $\text{H}_2\text{O}_2$ , but also is controlled by the concentration of  $\text{Fe}^{2+}$  and/or  $\text{Fe}^{3+}$  as catalyst [6,9]. The effects of initial  $\text{Fe}^{2+}$  concentrations on KN-R degradation at pH 3.0 are depicted in Fig. 6. Degradation rate undergoes a gradual acceleration when  $\text{Fe}^{2+}$  concentration increases from 0 to  $1.0 \text{ mmol/L}$ . It can be seen that the removal rate increased from 60% to 80% when the concentration of  $\text{Fe}^{2+}$  increased from 0 to  $0.75 \text{ mmol/L}$ . In the absence of  $\text{Fe}^{2+}$ , the main oxidant was the electro-generated  $\text{H}_2\text{O}_2$  which limited oxidation power. The enhanced decay should be attributed to the larger amount production of  $\text{HO}^\bullet$  in the medium with the increasing  $\text{Fe}^{2+}$ , however, with a further increase of  $\text{Fe}^{2+}$  concentration to  $1.0 \text{ mmol/L}$ , KN-R removal rate was declined. It was because that further increase in  $\text{Fe}^{2+}$  concentration will consume  $\text{HO}^\bullet$  with the following non-oxidizing reaction (equation (6)) [29]:



## 3.2.3. Change of UV-visible absorption bands

In order to clarify the changes of molecular and structural characteristics of KN-R at different electrolysis time in the electro-Fenton process, UV-visible spectra were measured in the course of oxidation.





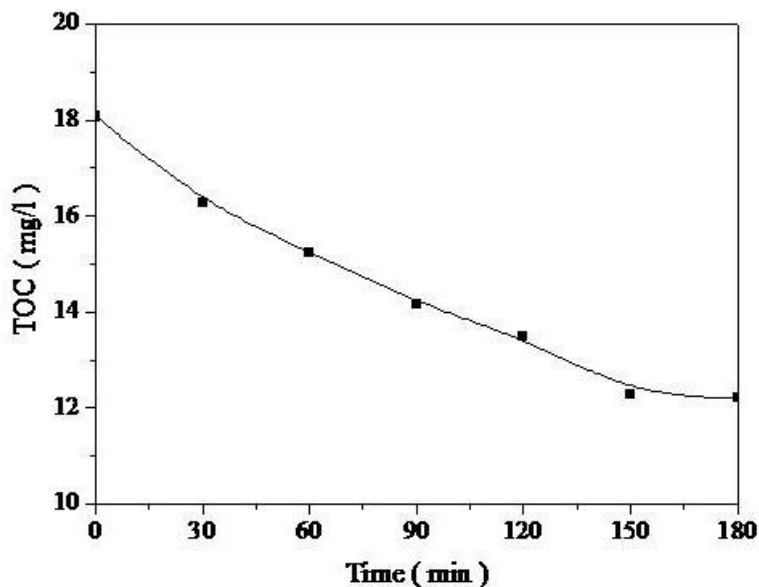
**Figure 7.** Change in the UV-visible spectra of KN-R with electrolysis time under the optimal electro-Fenton conditions:  $O_2$  flow rate of  $0.02\text{ m}^3/\text{h}$ ,  $0.05\text{ mol/L Na}_2\text{SO}_4$ , current density of  $2.0\text{ mA/cm}^2$ , pH 3.0  $0.75\text{ mmol/L Fe}^{2+}$  and  $M_{\text{graphite}}:M_{\text{GE}}:M_{\text{PTFE}} = 8:1:2$ .

As shown in Fig. 7, KN-R was characterized by one main band in the visible region with the absorption peak at ca. 590 nm, and the other two bands in the ultraviolet region with absorption peaks at around 310 and 240 nm, respectively. The peak at 590 nm might originate from the formation of hydrogen bond between amino-group and adjacent carbonyl groups in the matrix structure of anthraquinone, which strengthened the conjugate action between lone pair electrons and anthraquinone annulus. The weak absorption band at 310 nm was attributed to the transition of lone pair electrons in carbonyl groups. The characteristic absorption peak at 240 nm was ascribed to the matrix structure of anthraquinone, which embodied the existence of aromatic ring structure. With the electrolysis time extended, the absorption peak at 590 nm rapidly weakens accompanied with the disappearance of solution color, which indicates that the conjugated system of KN-R was destroyed completely. The two bands in the ultraviolet region also decreased due to the damage of aromatic ring structure.

#### 3.2.4. KN-R mineralization

Fig. 8 displays the mineralization of KN-R anthraquinone dye under the optimality conditions, i.e. pH = 3.0,  $0.75\text{ mmol/L}$  of  $\text{Fe}^{2+}$  concentration,  $2.0\text{ mA/cm}^2$  of current density and  $M_{\text{graphite}}:M_{\text{GE}}:M_{\text{PTFE}} = 8:1:2$ . As can be seen, maximum 33.3% TOC decay is obtained. This poor mineralization can be related to the formation of intermediate products such as short carboxylic acid and their complexes with iron ions, which are difficult to be oxidized completely by  $\text{HO}^\bullet$  generated in the medium [7,9,12]. But in the first 90 min a faster degradation is found, which can be accounted for the production of more  $\text{HO}^\bullet$  and that the major produced in this time period, aromatics intermediates can be more easily destructed [9]. The slower  $\text{Fe}^{2+}$  regeneration rate and the formation of stable  $\text{Fe}^{3+}$ -carboxylic acid complex, result in lower TOC decay at the second stage. These findings agree well

with the studies of Zhou et al. [5,6] and Brillas et al. [7,9,12] who also found the “platform behavior” during organics mineralization by electro-Fenton and photoelectro-Fenton processes.



**Figure 8.** TOC decay of KN-R under the optimal electro-Fenton conditions: O<sub>2</sub> flow rate of 0.02 m<sup>3</sup>/h, 0.05 mol/L Na<sub>2</sub>SO<sub>4</sub>, current density of 2.0 mA/cm<sup>2</sup>, pH 3.0 0.75 mmol/L Fe<sup>2+</sup> and M<sub>graphite</sub>:M<sub>GE</sub>:M<sub>PTFE</sub> = 8:1:2.

#### 4. CONCLUSIONS

In this study, GE/graphite-PTFE GDE was prepared and the efficiency of electro-generated H<sub>2</sub>O<sub>2</sub> was evaluated by an undivided electro-Fenton system. The amount of electro-generated H<sub>2</sub>O<sub>2</sub> using GE/graphite-PTFE GDE (M<sub>graphite</sub>:M<sub>GE</sub>:M<sub>PTFE</sub> = 8:1:2) was nearly three times higher than that of graphite-PTFE GDE (M<sub>graphite</sub>:M<sub>PTFE</sub> = 8:2) due to the accelerated electron transfer rate resulting from GE. The highest H<sub>2</sub>O<sub>2</sub> concentration was obtained at pH 3.0 and the current density of 2.0 mA/cm<sup>2</sup>. KN-R could be degraded quickly and efficiently by the proposed electro-Fenton process using GE/graphite-PTFE GDE. Under the pH 3.0 and 0.75 mmol/L Fe<sup>2+</sup> concentration, ca. 80% KN-R with the initial concentration of 80 mg/L could be degraded and 33.3% mineralization rate was achieved after 180 min electrolysis.

#### ACKNOWLEDGMENTS

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#### References

1. M.A. Brown and S.C. De Vito, *Crit. Rev. Environ. Sci. Technol.*, 23 (1993) 249
2. V. Meshko, L. Markovska, M. Mincheva, and A.E. Rodrigues, *Water Res.*, 35 (2001) 3357

3. C. Galindo, P. Jacques, and A. Kalt, *Chemosphere*, 45 (2001) 997
4. M.C. Venceslau, S. Tom, and J.J. Simon, *Environ. Technol.*, 15 (1994) 917
5. M.H. Zhou, Q.H. Yu, and L.C. Lei, *Dyes Pigments*, 77 (2008) 129
6. M.H. Zhou, Q.H. Yu, L.C. Lei, and G. Barton, *Sep. Purif. Technol.*, 57 (2007) 380
7. E. Isarain-Chavez, R.M. Rodriguez, J.A. Garrido, C. Arias, F. Centellas, P.L. Cabot, and E. Brillas, *Electrochim. Acta*, 56 (2010) 215
8. Z.H. Ai, T. Mei, J. Liu, L.Z. Zhang, K.J. Deng, and J.R. Qiu, *J. Phys. Chem. C*, 112 (2008) 11929
9. C. Flox, S. Ammar, C. Arias, E. Brillas, A.V. Vargas-Zavala, and R. Abdelhedi, *Appl. Catal. B: Environ.*, 67 (2006) 93
10. Ch. Comninellis, *Electrochim. Acta* 39 (1994) 1857
11. E. Brillas, R. Sauleda, and J. Casado, *J. Electrochem. Soc.*, 145 (1998) 759
12. S. Randazzo, O. Scialdone, and E. Brillas, *J. Hazard. Mater.*, 193 (2011) 1555
13. A. Wang, J. Qu, H. Liu, and J. Ru, *Appl. Catal. B: Environ.*, 84 (2008) 393
14. Y.R. Wang and W. Chu, *Water Res.*, 45 (2011) 3883
15. J.S. Do and C.P. Chen, *J. Appl. Electrochem.*, 24 (1994) 936
16. S. Murugesan, K. Myers, and V.R. Subramanian, *Appl. Catal. B: Environ.*, 103 (2011) 266
17. A. Dhaouadi, L. Monser, and N. Adhoum, *Electrochim. Acta*, 54 (2009) 4473
18. N. Daneshvar, S. Aber, V. Vatanpour, and M.H. Rasoulifard, *J. Electroanal. Chem.*, 615 (2008) 165
19. S. Hammami, N. Oturan, N. Bellakhal, M. Dachraoui, and M.A. Oturan, *J. Electroanal. Chem.*, 610 (2007) 75
20. A.A. Gallegos and D. Pletcher, *Electrochim. Acta*, 44 (1999) 2483
21. C.P. Leon and D. Pletcher, *J. Appl. Electrochem.*, 25 (1995) 307
22. A.K. Geim, *Science* 24 (2009) 1530
23. A.K. Geim, K.S. Novoselov, S. V. Morozov, D. Jiang, Y. Zhang, and S. V. Dubonos, *Science* 306 (2004) 666
24. W.S. Hummers and R.E. Offeman, *J. Am. Chem. Soc.*, 80(1958) 1339
25. N.I. Kovtyukhova, P.J. Ollivier, B.R. Martin, T.E. Mallouk, S.A. Chizhik, and E.V. Buzaneva, *Chem. Mater.*, 11 (1999) 771
26. A. Özcan, Y. Sahin, and M.A. Oturan, *Chemosphere* 73 (2008) 737
27. B.G. Kwon, D.S. Lee, and N. Kang, *Water Res.*, 33 (1999) 2110
28. L. Szpyrkowicz, C. Juzzolino, and S.N. Kaul, *Water Res.*, 35 (2001) 2129
29. M. Panizza and G. Cerisola, *Water Res.*, 35(2001) 3987